

Accepted Manuscript

Using a detailed inventory of a large wastewater treatment plant to estimate the relative importance of construction to the overall environmental impacts

Serni Morera, Lluís Corominas, Miquel Rigola, Manel Poch, Joaquim Comas



PII: S0043-1354(17)30440-2

DOI: [10.1016/j.watres.2017.05.069](https://doi.org/10.1016/j.watres.2017.05.069)

Reference: WR 12948

To appear in: *Water Research*

Received Date: 21 December 2016

Revised Date: 23 May 2017

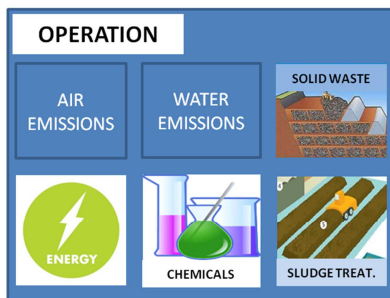
Accepted Date: 29 May 2017

Please cite this article as: Morera, S., Corominas, Lluí., Rigola, M., Poch, M., Comas, J., Using a detailed inventory of a large wastewater treatment plant to estimate the relative importance of construction to the overall environmental impacts, *Water Research* (2017), doi: 10.1016/j.watres.2017.05.069.

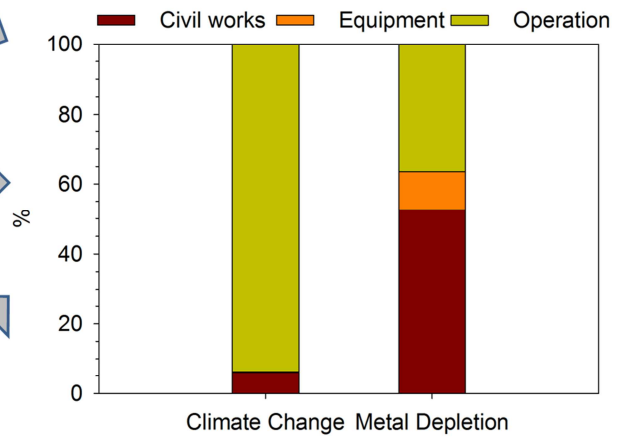
This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2017. This manuscript version is made available under the CC-BY-NC-ND 4.0 license <http://creativecommons.org/licenses/by-nc-nd/4.0/>





LCA of a large WWTP



1 Using a detailed inventory of a large wastewater treatment plant to
2 estimate the relative importance of construction to the overall
3 environmental impacts

4 Serni Morera¹, Lluís Corominas², Miquel Rigola¹, Manel Poch¹, Joaquim Comas^{1,2}

5 ¹ LEQUIA, Institute of the Environment. University of Girona. Campus Montilivi. Carrer
6 Maria Aurèlia Capmany, 69, E-17003 Girona. Catalonia. Spain

7 ² ICRA, Catalan Institute for Water Research, Scientific and Technological Park of the
8 University of Girona, Emili Grahit, 101, E-17003 Girona, Spain

9

10 Corresponding author: jcomas@icra.cat

11 Telephone: 0034 972 183380

12

13 **Abstract**

14 The aim of this work is to quantify the relative contribution to the overall environmental
15 impact of the construction phase compared to the operational phase for a large conventional
16 activated sludge wastewater treatment plant (WWTP). To estimate these environmental
17 impacts, a systematic procedure was designed to obtain the detailed Life Cycle Inventories
18 (LCI) for civil works and equipment, taking as starting point the construction project budget
19 and the list of equipment installed at the Girona WWTP, which are the most reliable
20 information sources of materials and resources used during the construction phase. A detailed
21 inventory is conducted by including 45 materials for civil works and 1240 devices for the
22 equipment. For most of the impact categories and different life spans of the WWTP, the
23 contribution of the construction phase to the overall burden is higher than 5% and, especially
24 for metal depletion, the impact of construction reaches 63%. When comparing to the WWTP
25 inventories available in Ecoinvent the share of construction obtained in this work is about 3
26 times smaller for climate change and twice higher for metal depletion. Concrete and
27 reinforcing steel are the materials with the highest contribution to the civil works phase and
28 motors, pumps and mobile and transport equipment are also key equipment to consider during
29 life cycle inventories of WWTPs. Additional robust inventories for similar WWTP can
30 leverage this work by applying the factors (kg of materials and energy per m³ of treated
31 water) and guidance provided.

32 **Keywords:** Civil works, equipment, LCA, life span, WWTP, detailed inventory

33

34

35 1. INTRODUCTION

36 Wastewater treatment plants (WWTPs) are designed to reduce the impact of wastewater
37 generated in urban systems before discharging it to the receiving water bodies. Despite their
38 beneficial contribution to the environment, they also generate environmental impacts during
39 their construction, operation and dismantling. Life Cycle Assessment (LCA) is the most
40 commonly used methodology to evaluate the global environmental impacts of WWTPs. All
41 impacts produced throughout a WWTPs' lifetime, from their construction and operation until
42 deposition or recycling, are included in the assessment, however not all are included
43 systematically.

44
45 This paper focuses on the construction phase, including both civil works and equipment, as
46 we believe that it is a stage which has been understudied (Remy and Jekel, 2008). As an
47 indication, in the review of LCA application to WWTPs from Corominas et al. (2013) only 22
48 studies (out of 45 reviewed) included the construction phase and only 15 provided their own
49 inventories. Some studies report a contribution of construction to the overall WWTP impacts
50 lower than 5% (e.g. Emmerson et al., 1995; Vlasopoulos et al., 2006). Besides, coming up
51 with detailed own inventories is tedious and time consuming. This persuades some
52 researchers/practitioners to perform detailed inventories. A potential solution (as applied in
53 Foley et al., 2010) is the estimation of the volume of reinforced concrete which is then used as
54 a multiplier for the estimation of other construction phase materials taking as a reference the
55 inventories provided by Ecoinvent (Doka et al., 2007). In the case that construction is
56 included, a minority of studies include equipment with limited information (e.g., Tillman et
57 al., 1998; Lundin et al., 2000; Machado et al., 2007; Ortiz et al., 2007; Stokes et al., 2010;
58 Foley et al., 2010; Risch et al., 2015), the rest only take into account civil works. In general,

59 data on materials use (concrete, steel, etc.) can be collected but information on the utilities
60 (e.g. energy consumption of constructing vehicles) is missing. In spite of these difficulties,
61 according to us, we should account for both civil works (production of materials, transports
62 from factory to workplace and combustible consumed) and also equipment (i.e. thousands of
63 devices, including diffusers, pumps and blowers), which has to be replaced several times
64 during the life span of the WWTP. The low level of detail of construction inventories
65 published so far and the large number of assumptions normally taken questions the validity of
66 the estimates of the share of construction to the overall environmental impacts of WWTPs.

67
68 Hence, the objective of this work is to estimate the relative importance of construction to the
69 overall environmental impacts of a large WWTP thanks to the development of a detailed
70 inventory. The main novelty compared to previous published work lies in the provision of
71 detailed inventories for civil works (including 45 different materials) and equipment (1240
72 devices), operation and dismantling. The inventory is modular as the WWTP has been divided
73 into 5 units: pumping & pretreatment, primary treatment, secondary treatment, sludge line and
74 deposition, and buildings and services (offices building and exterior landscaping).

75
76 The paper is organized as follows. First, in the methodology section, a systematic procedure
77 to obtain detailed Life Cycle Inventories for civil works and equipment is introduced. Then
78 this methodology is used to obtain the construction inventories for a real WWTP. In the
79 results and discussion section, first, the contribution of the WWTP construction (civil works
80 and equipment) and operation to the overall impact are compared, including a sensitivity
81 analysis to evaluate the relevance of the WWTP life span as well as an uncertainty analysis of
82 the inventories of the civil works, equipment and operation. Second, the contribution of the

83 construction (civil works and equipment), including also its dismantling, and operation to the
84 impact for each operational unit and impact category is analyzed. Third, a cumulative effect to
85 the overall impact of the civil works, equipment and operation for each operational unit and
86 resources consumed (material, energy, transport) is shown. Finally, a set of simplifications in
87 the construction inventory are evaluated to provide guidance to LCA practitioners.

88

89 **2. METHODS**

90 **2.1. Procedure followed to obtain a detailed construction inventory**

91 Detailed inventories for civil works includes the following steps (Figure 1): (1) Obtain the
92 construction budget for the WWTP. In the budget, all elements needed for civil works (e.g.,
93 excavations, handrails, concrete, etc.) are listed, along with the price and the amount of each
94 one. (2) Then, all these elements in the budget are identified and grouped in a simplified list
95 (e.g., excavation of a representative type of soil, a representative type of concrete, etc.). (3)
96 Once all elements of the budget are identified and grouped, it is necessary to search for
97 equivalent elements in a specialized constructive database of reference (in our case the local
98 database Banc BEDEC, which is used by constructors to make their budgets). This database
99 provides all the necessary information about materials and energy consumed to build a unit of
100 each element. (4) Finally, the material and energy inventories are calculated by relating the
101 elements obtained from the construction budget to the equivalent elements of the reference
102 database.

103 (FIGURE 1)

104 The procedure for equipment is as follows: (1) Obtain the list of equipment from the
105 equipment maintenance plan or, if it is not available, using as-built documents. (2) The
106 equipment is characterized and organized in classes. (3) Obtain the environmental product
107 declarations (EPDs) of each class of equipment. In the EPDs, the material and energy needed
108 for the manufacturing of each equipment are recorded. Sometimes, specific EPDs are not
109 available, in which case it is necessary to deduce information from similar elements of each
110 group. (4) Obtain the inventory, relating the information from the EPDs with the list obtained
111 in the classification of the equipment.

112 Five WWTP units were considered in the case study: (1) pumping + pretreatment, (2) primary
113 treatment, (3) secondary treatment, (4) sludge line and deposition, and (5) buildings and
114 services. Table 1 provides a description of the elementary processes considered for each
115 operational unit considered in this work. The procedure was applied to obtain detailed
116 inventories for each one of these units of the WWTP.

117 (TABLE 1)

118 **2.2. Life cycle assessment**

119 The environmental assessment was conducted following the ISO 14040 and 14044 standards
120 (ISO 14040, 2006; ISO 14044, 2006) which define four stages: (1) goal and scope definition;
121 (2) inventory analysis; (3) environmental impact assessment; and (4) interpretation.

122 *2.2.1. Goal and scope definition*

123 The goal was to perform an LCA of the Girona WWTP, considering the construction of the
124 plant (civil works + equipment), the operation and dismantling, and making an individual

125 analysis for the five WWTP units. The functional unit of the study is 1 m³ of treated
126 wastewater assuming that the WWTP is working at full capacity. In fact, WWTPs are
127 designed and constructed to serve a specific capacity. However, the treated flow does not
128 always match with the design flow. We assumed the lifetime of the WWTP to be 20 years
129 (Renou et al., 2008) whereas the lifetime of the equipment was 15 years (Lundin et al., 2000).
130 The sensitivity of the lifetime of the WWTP on the LCA assessment is addressed in a
131 sensitivity analysis. Savage values for the equipment are considered when the life spans of the
132 WWTP and of the equipment do not match.

133 The studied WWTP is located in Girona (Catalonia, NE of Spain). It treats the wastewater
134 from the main city and different nearby towns located around the WWTP, before the effluent
135 is discharged into the Ter River. The plant has a capacity of 206,250 population equivalents
136 (PE), which corresponds to a design flow rate of 55,000 m³·d⁻¹. However, in the year 2013 the
137 WWTP of Girona treated on average 42,000 m³·d⁻¹. The water line consists of a Modified
138 Ludzack-Ettinger (MLE) configuration with biological removal of organic matter and
139 nitrogen and chemical removal of phosphorus. The sludge line consists of thickening,
140 anaerobic sludge digestion with electricity production from the biogas, and sludge
141 dewatering. The dewatered sludge is sent to a nearby composting plant. The composting plant
142 is a private installation located 20 km away from the WWTP, and it treats not only the sludge
143 from the WWTP but also other organic residues from other facilities. Figure 2 shows a
144 scheme of the WWTP with a separation of each analyzed operational unit and an indication of
145 the operational data used to perform the LCA of the operation.

146 (FIGURE 2)

147 Pumping and pretreatment includes five pumps, two channels that are 16 meters long, two
148 sieves, screening and grease separation. Primary treatment has three primary settlers, each
149 with a total capacity of 412 m³. Secondary treatment consists of three different reactors with a
150 total capacity of 29,620 m³ and three settlers, each with a capacity of 5,027 m³. The sludge
151 line includes two thickeners of 16 m of diameter, two primary digesters with a 3,432 m³
152 volume, two secondary digesters with an 814 m³ volume, two dewatering devices and a
153 cogeneration device.

154 The system boundaries (Figure 3) consider the WWTP's construction, operation and
155 dismantling. The production of all materials (and their transport) and energy used to build the
156 WWTP (civil works + equipment) are accounted. The dismantling is considered in this study
157 for the most abundant materials for both civil works and equipment, assuming that concrete
158 and reinforced concrete are disposed at a landfill for inert waste, 91% of metals are recycled
159 (Sansom and Avery, 2014) and 25% of plastics are recycled, 34% of plastics are incinerated
160 with electricity recovery and the rest disposed in a landfill (Plastics Europe, 2012). Operation
161 includes the electricity and chemicals consumed, the gases and water emissions from the
162 WWTP to the environment, the deposition of residues from the pretreatment and primary
163 treatment of the WWTP (solid residues, sand and greases), the sludge composting (but not its
164 application in agriculture) and the electricity produced. Transports from suppliers to the
165 WWTP during the construction and operation phases are considered as well. The construction
166 of the composting plant is not within the system boundaries.

167 (FIGURE 3)

168

169 2.2.2. *Inventory analysis*

170 The procedure described in section 2.1. has been applied to obtain the construction
171 inventories of each of the five WWTP units in terms of materials and energy (Table 2).
172 Hence, for all WWTP stages (pre-treatment, primary treatment, secondary treatment, sludge
173 line, buildings and services) we have accounted for materials and equipment. A summary of
174 the construction inventory can be found in Table 2. 45 different types of materials have been
175 used in the construction of the WWTP of Girona (see Table S-1 from supporting information
176 for a complete material inventory). The WWTP has 1240 devices, including large equipment
177 (e.g. blowers, pumps, motors, mixers, heat exchangers, compressors, diffusers) and small
178 equipment (e.g. valves, gates, probes). Tables S-2a and S-2b lists all the devices and
179 corresponding materials for all equipment installed per operational unit while section 1 in
180 Supplementary Information details how inventories have been performed for motors and
181 pumps. Tables S-3 and S-4 and section 1 in supplementary information provide independent
182 inventories for all pumps and motors typically existing in WWTPs (which can be directly
183 applied to other studies). Transports have been estimated considering the weight of the
184 materials and assuming an average distance of 40 km.

185 Table 2 also includes the inventory of the operational phase. That information has been
186 provided by Trargisa S.A., the company that manages the Girona WWTP. The chemicals used
187 in the WWTP are iron chloride, sodium aluminate, polyelectrolyte, antifoaming and
188 antioxidant. For iron chloride the correspondence in the Ecoinvent database has been used.
189 For sodium aluminate no equivalent has been found in Ecoinvent, and hence we have created
190 our own process using technical information on the chemical production. For chemicals with
191 no technical information available an alternative similar chemical has been selected from the

192 Ecoinvent database. Acrylonitrile is the proxy for polyelectrolyte and silicon product is the
193 proxy for antifoaming and antioxidant. The composting process used in this assessment is the
194 one described by Remy (2010), which provides inventory data tailored to current composting
195 processes (more recent than the processes available in the Ecoinvent database). The direct
196 emissions of greenhouse gases from secondary treatment, from biogas combustion and from
197 the degradation of the organic matter and nutrients emitted to the river, have been estimated
198 applying the factors from Foley et al. (2010), which are 0.01 kg N₂O-N/kg N denitrified in
199 secondary treatment, 16.02 gCH₄/Nm³ biogas and 0.73 g N₂O/Nm³ biogas from biogas
200 combustion, and 0.025 kg CH₄/kg COD discharged to the river and 0.0025 kg N₂O-N/kg N
201 discharged to the river.

202 (TABLE 2)

203 The factors provided in Table 2 (kg of materials and energy per m³ of treated water) can be
204 reused to develop robust inventories for the construction, operation and dismantling of
205 treatment works similar to the Girona WWTP.

206 2.2.3. *Environmental Impact Assessment*

207 The types of materials, energy sources and emissions from the inventories have been matched
208 to their corresponding equivalents in the Ecoinvent 3 database (Weidema et al., 2013). The
209 potential environmental impacts have been calculated through the use of LCIA
210 characterization factors related to a sub-set of impact categories from ReCiPe (H) 1.13
211 (Goedkoop et al., 2013). We have included climate change (CC), ozone depletion (OD),
212 freshwater eutrophication (FE), marine eutrophication (ME), human toxicity (HT), metal

213 depletion (MD) and fossil depletion (FD). All inventories used for the materials and energy
214 production processes in this study have been taken from Ecoinvent 3 (Weidema et al., 2013).
215 The assessment has been conducted using SimaPro 8.3.0.0.

216 **2.3. Sensitivity and uncertainty analysis**

217 One of the most influential factors in the LCA of WWTPs is the selection of the life span
218 (Risch et al., 2015). A sensitivity analysis has been conducted to evaluate the influence of the
219 selection of the life span of the WWTP to the environmental impacts. Hence, besides the base
220 case of 20 years of life span, we have evaluated as well 10 and 30 years life span.

221 Uncertainty analysis has been applied to analyze the variability of results due to the variation
222 of uncertain input parameters over their whole domain of uncertainty, using the Monte-Carlo
223 engine in Simapro. We have assumed an uncertainty of 13% for civil works which
224 corresponds to the “unexpected issues” item in the execution budget of the Girona WWTP.
225 This 13% covers the uncertainty related to the execution of the project given a certain budget,
226 and does not account for possible sources of variability that one could find in other case-
227 studies. Uncertainty has been only assessed on the amount of material used for the machinery
228 with the highest influence on the impact (i.e. pumps and motors). The uncertainty range has
229 been defined separately for pumps and motors after compiling EPDs from different vendors
230 and conducting a linear regression between the usage of materials and the power (see section
231 2 and Tables S-5 and S-6 of supplementary information for further details on the uncertainty
232 analysis for motors and pumps). The error associated to those regressions is assumed to be the
233 uncertainty, ranging from 10% (for large equipment) to 400% (for small equipment).

234 With regards to the uncertainty associated to the WWTP operation, it was addressed after
235 evaluating the variability in the concentrations of pollutants at the influent of the WWTP, the

236 chemicals consumption and the electricity consumption over three consecutive years. The
237 uncertainty values were obtained calculating the mean and the maximum and minimum
238 values of the values. In addition, we have used the factor reported in Foley et al. (2010) to
239 include N₂O emissions (0.0003 to 0.03 kg N₂O-N/kg N denitrified in the secondary
240 treatment). Uniform probability distribution functions have been applied to all sources of
241 uncertainty evaluated.

242 **3. RESULTS AND DISCUSSION**

243 **3.1. Environmental impact assessment of the construction and operation of the entire** 244 **plant**

245 When assuming a life span of the WWTP of 20 years (Figure 4, middle bar for each impact
246 category), the contribution of the construction (civil works + equipment) to the overall burden
247 is higher than 5% for most environmental impact categories evaluated. For MD the
248 contribution of construction is as large as 63%, with a share of approximately 52% for civil
249 works (i. e. $2.00 \cdot 10^6$ kg, mainly reinforcing steel) and of 10% for the equipment (e.g., $4.1 \cdot 10^4$
250 kg of cast iron and $1.5 \cdot 10^4$ kg of steel, without considering the replacement). For HT and FD,
251 the contribution of construction to the overall impact is approximately 16%. In the case of FD
252 most of the construction burden comes from the production of materials (only 3% of FD
253 relates to energy consumed during the construction phase). For CC, OD and FE, the
254 contribution of construction represents between 5 and 10% of the overall impact. It is worth
255 mentioning that construction contributes to FE through emissions generated during the
256 production of materials, such as steel and concrete as well as through landfill emissions. Only
257 for ME the contribution to the impact of civil works and equipment is less than 1%. Keeping
258 in mind that the cut-off criteria defined in the LCA Handbook (European Commission, 2010) is

259 5% influence on the environmental impacts it is clear that one cannot systematically omit the
260 construction phase in LCA studies for WWTPs of similar configuration and size of the one
261 studied in this paper. It is worth mentioning that the inclusion of the dismantling/end-of-life
262 of the WWTP implies recycling most of metals from the equipment. Recycling reinforcing
263 steel would decrease the MD burden down to negligible values for construction. However, the
264 latter is not a common practice due to the large costs involved in the separation of concrete
265 and the reinforcing steel.

266 (FIGURE 4)

267 The results for the sensitivity analysis of the WWTP life span shows little effect on the
268 categories CC, OD, FE, ME and FD. For HT and MD, the longer the life span is the larger the
269 contribution of equipment to the overall environmental impacts, as replacement will be
270 needed. In contrast, when the life span considered is 30 years, the contribution of civil works
271 significantly decreases for the MD category.

272 Overall, even when accounting for uncertainty, the conclusion that construction cannot be
273 neglected remains valid. The uncertainty related to the operation phase has a larger influence
274 on most of the environmental impacts than the one related to the construction phase (Figure
275 5). The propagation of uncertainty of the emission factor of N₂O emissions has a large impact
276 on climate change (see errors bars for operation in CC, Figure 5). The large uncertainty
277 behind fugitive GHG emissions can change the relative contribution of construction to the
278 overall impacts. If fugitive GHG emissions would not be considered the contribution of
279 construction to climate change would be higher (close to 10%). The climate change results

280 vary by 5% when propagating the uncertainty from the operation. In the case of MD, the
281 propagation of uncertainty comes equally from civil works, equipment and operation.
282 (FIGURE 5)

283

ACCEPTED MANUSCRIPT

284 **3.2. The contribution of construction to the environmental burden found in the**
285 **literature**

286 Figure 6 shows the results of the contribution of construction and operation to the global
287 environmental impacts for relevant papers in literature which studied a similar wastewater
288 treatment technology and specified the life span of infrastructure. Overall, the contribution of
289 construction ranges from 1% to 63%. While operation and civil works were considered in all
290 studies presented in Figure 6 (only a few included GHG emissions), equipment was only
291 considered in Vlasopoulos et al. (2006), Machado et al. (2007), Ortiz et al. (2007) and Renou
292 et al. (2008).

293 (FIGURE 6)

294 Previous studies reported a share of construction to the overall impacts lower than 5%
295 (Emmerson et al., 1995; Vlasopoulos et al., 2006). We understand that there is probably an
296 underestimation of concrete (we estimated that 3 times more concrete would be needed in a
297 WWTP of the same size compared to Emmerson et al., 1995), and that no databases of
298 construction elements such as Banc BEDEC have been used. The construction inventory for
299 Vlasopoulos et al. (2006) was collected by questionnaire contact to over 160 equipment
300 designers, manufacturers and suppliers, but not based on real information from existing
301 systems and following a detailed budget. No detailed information was provided on how the
302 construction inventory was executed and on the assumptions made.

303 Renou et al. (2008) concluded that the construction of an activated sludge plant had a
304 contribution of 11% to the CC impact category, close to the 8% reported in this study.
305 However, for the HT impact category they reported a share of construction of 1%, which is

306 much lower than the 18% obtained in this study. The difference might be explained by the
307 fact that they did not include equipment in their inventories.

308 Some studies reported a share of construction to the overall impacts between 20 and 30%
309 (Machado et al., 2007; Ortiz et al., 2007). The difference compared to our study might be
310 explained by the shorter life span they applied and because of the economy of scale. In
311 addition, the inclusion of N₂O and CH₄ emissions in the operation might have a large
312 influence downgrade the contribution of construction to the overall impacts.

313 Using the inventory values reported in Foley et al. (2010) (for their scenario 4Bii) and running
314 ReCiPe (H) 1.13 we obtained a contribution for all categories lower than 5% for the
315 construction, except for MD where the contribution of construction was of 18%. From the
316 interpretation of that inventory (it is not explained in detail) we understand that in Foley et al.
317 (2010) they considered materials to build mainly the secondary treatment, but did not include
318 in detail all possible stages (pumping, pretreatment, primary treatment and sludge line).
319 Whereas in Foley et al. (2010) they applied a factor of 77 kg of steel per m³ of concrete in this
320 study a factor of 90 kg of steel per m³ was used. That increases the amount of metals, and
321 hence has an effect specially for HT and MD. In addition, whereas in Foley et al. (2010) they
322 accounted for the production of two materials (steel and copper) and the mass for 3 type of
323 equipment (pumps, motors and blowers), we included 30 additional materials and 1240
324 devices (Table S-2a).

325 Risch et al. (2015) reports a contribution of 20% for CC, higher than our study (around 8%), a
326 similar one for HT (18%) and a lower contribution for OD, FE, but specially for MD and FD,
327 even though Risch et al. (2015) use a life span of 30 years. The exhaustive inventory for the

328 equipment and civil works phases in our study can be the explanation for these significant
329 differences.

330 Ecoinvent provides some inventories for construction and operation for 30 years lifespan of 5
331 different WWTPs. After recalculating the results for 20 years life span for the 5 WWTPs, the
332 average of the construction and operation impacts shows a contribution of the construction for
333 CC (32%) and OD (29 %), which are values higher than our results (8% and 10%
334 respectively). Similar results are obtained for ME, HT and FD, with a contribution of the
335 construction of 0%, 28% and 24% of each one. Finally, for FE the contribution is very similar
336 to our study (around 8%) while for MD the contribution is lower (45% against 65% obtained
337 in our study). Apparently there is a disagreement between the results reported in Ecoinvent
338 and other studies. That disagreement might come from an overestimation of the amount of
339 concrete. Whereas in Ecoinvent a factor of 6268 kg/m^3 is applied, in this study we obtained a
340 factor of 927 kg/m^3 . The latter, is within the order of magnitude of other studies, such as 579
341 kg/m^3 in Foley et al. (2010) or 428 kg/m^3 in Emmerson et al. (1995).

342 **3.3. Identification of key resources/processes that contribute the most to the** 343 **environmental burden**

344 In this section we identify which are the key resources/processes in the construction and
345 operation of the WWTP of Girona. A summary of the contribution of each of the treatment
346 units to construction, equipment and operation related impacts is provided in Table 3 and
347 illustrated in Figure 7. With regards to civil works, the production of concrete used in the
348 construction of the secondary treatment (mainly the biological reactors) and the sludge line
349 (mainly the anaerobic digester) is the process contributing the most to CC impacts. Then, the
350 production of reinforcing steel (also used to build the biological reactors, the digester and

351 thickeners) is the process that contributes the most to MD (with a contribution on its own of
352 41% of the total impact, 25% from the biological reactors and 11% from the digester) and to
353 HT (a contribution of 7.6% of the total). Recycling the steel from the reinforced concrete
354 would significantly decrease the MD and HT impacts, although probably at expenses of
355 increase in CC due to the usage of machinery. However, this is not a common practice due to
356 high costs. There is a slight contribution of the production of plastics to HT and MD (lower
357 than 4%). The equipment contributes to 11% of the MD, 3.4% of the HT, and is almost
358 negligible for the CC impacts. Metals consumed during the production of equipment,
359 including pumps and motors, transport and mobile equipment, are contributing the most to
360 MD. When it comes to operation, and looking into CC impact category, then the electricity
361 consumed for pumping & pretreatment and secondary treatment (aeration) correspond to
362 about 20% of the total impact, and the direct GHG emissions emitted from the biological
363 reactor (35% of the total impact, but with large associated uncertainty) are the largest
364 contributors. The electricity produced at the treatment plant through cogeneration corresponds
365 to an avoided impact of about 10%. When it comes to HT and MD, then the chemicals used in
366 primary treatment (i.e. iron chloride) and secondary treatment (i.e. sodium aluminium) also
367 play a significant role, as they contribute to 30% of HT and 20% of MD. Finally, the
368 deposition of solids from the pretreatment to landfill contribute to 17% of the HT. The
369 avoided electricity production has also a positive impact on the HT (16%).

370 Additionally, Figures S-1 to Figure S-4 illustrate the contribution of civil works, equipment
371 and operation for each operational unit for the other studied impact categories (OD, FE, ME
372 and FD, respectively). In Figures S-2 and S-3 the high influence of the discharge of nitrogen
373 and phosphorus from the WWTP is observed for the eutrophication impact categories.

374 Previous published works already identified that electricity consumption, chemicals and
375 primary solids are the main contributors to CC, MD and HT (e.g. Hospido et al., 2004,
376 Rodriguez-Garcia et al., 2011) and nutrients discharged to the receiving water bodies
377 (assigned to secondary treatment in this study) are the most important contributors to
378 eutrophication (e.g. Hospido et al., 2008). The fact that concrete and reinforcing steel are also
379 important contributors to the CC and MD impacts has been reported before in Machado et al.,
380 (2007). The significant contribution of equipment to MD is an important highlight of this
381 study since has never been reported before, and as estimated in this study can contribute up to
382 11% of the MD impacts.

383 (FIGURE 7)

384 **3.4. Influence of inventory simplifications on the environmental impacts**

385 As an exercise of simplification we took the key resources/processes identified in previous
386 sections and calculated the overall coverage of the estimation of the environmental impacts
387 based on factors for materials and energy provided in Table 2. An LCA exercise which would
388 include concrete, reinforcing steel and plastics, together with the operational data (energy,
389 GHG emissions, chemicals and deposition of solids from the pretreatment), would allow to
390 estimate between 90 and 99% of the impact in CC, OD, FE, ME and FD categories. In order
391 to encompass a minimum 90% of the HT and MD impacts, metals consumed during the
392 production of pumps, motors, transport and mobile equipment should be considered in
393 addition to the data mentioned before. From our study, we can conclude that buildings and
394 service units can be omitted from the analysis.

395 At another simplification level, the inclusion of civil works, equipment and operation of
396 secondary treatment (i.e. biological reactor and settler) and sludge line (i.e. thickeners,

397 anaerobic digester and centrifuges) would result in about 75% coverage of the total impact for
398 the different impact categories in this study.

399 It is also important to mention that these conclusions are based in a single application of this
400 approach in an activated sludge WWTP of 206,250 PE. For this reason, further work should
401 be conducted to analyze different treatment technologies and different sizes to draw more
402 general recommendations.

403

404 **4. CONCLUSIONS**

405 An inadequate level of detail for materials and resources inventoried during construction,
406 operation and dismantling processes (e.g. considering or not GHG or metals for the
407 equipment) significantly influences the estimation of the share of construction to the global
408 environmental impacts of WWTPs. With our case study, we demonstrate that for most of the
409 impact categories the contribution of construction to the overall burden is higher than 5% and
410 especially for metal depletion the impact of civil works plus equipment reaches 63%. When
411 comparing to the WWTP inventories available in Ecoinvent, the share of construction is about
412 three times smaller for climate change but twice higher for metal depletion. Although the
413 equipment has a smaller impact than civil works, it must be considered as well; depending on
414 the category and the number of times it has to be replaced, its impact cannot be neglected,
415 particularly in the HT and MD impact categories. Concrete and reinforcing steel are the
416 materials with the highest contribution to the civil works phase and motors and pumps are
417 also key equipment to consider during life cycle inventories of WWTPs. The factors obtained
418 with the most detailed inventory published so far facilitates the development of robust
419 inventories for WWTPs construction, operation and dismantling.

420 5. ACKNOWLEDGMENTS

421 The authors would like to thank Trargisa S.A. for providing the data and feedback on the
422 WWTP of Girona, Voltes S.L.U. for the technical support, Banc BEDEC, the Ministry of
423 Economy and competitiveness for the Ramon and Cajal grant from Lluís Corominas (RYC-
424 2013-14595) and the REaCH project (CTM2015-66892-R). Serni Morera's FI scholarship
425 (2015FI_B2 00071), STSMscholarship from EU-Cost Action Water2020 (ECOST-STSM-
426 ES1202-110614-044095). LEQUIA and ICRA were recognized as consolidated research
427 groups by the Catalan Government with codes 2014-SGR-1168 and 2014-SGR-291,
428 respectively.

429

430 6. REFERENCES

- 431 Corominas, L., Foley, J., Guest, J. S., Hospido, A., Larsen, H. F., Morera, S., Shaw, A.,
432 (2013). Life cycle assessment applied to wastewater treatment: state of the art. *Water*
433 *Res.*, 47 (15), 5480–92.
- 434 Doka, G., (2007). Life Cycle Inventories of Waste Treatment Services. Ecoinvent report No.
435 13, Swiss Centre for Life Cycle Inventories, Dübendorf, December 2007.
- 436 Emmerson, R. H. C., Morse, G. K., Lester, J. N., Edge, D. R., (1995). The Life-Cycle
437 Analysis of Small-Scale Sewage-Treatment Processes. *Journal of CIWEM*, 9, 317–325.
- 438 European Commission – Joint Research Centre – Institute for Environment and
439 Sustainability: International Reference Life Cycle Data System (*ILCD*) *Handbook -*
440 *General guide for Life Cycle Assessment - Detailed guidance*. First edition March 2010.
441 EUR 24708 EN. Luxembourg. Publication Office of the European Union; 2010.

- 442 Foley, J., De Haas, D., Hartley, K., Lant, P., (2010). Comprehensive life cycle inventories of
443 alternative wastewater treatment systems. *Water Res.*, 44 (5), 1654–66.
- 444 Frischknecht, R., Jungbluth, N., Hans-Jörg, A., Doka, G., Dones, R., Heck, T., Hellweg, S.,
445 Hischer, R., Nemecek, T., Rebitzer, G., Spielmann, M., (2005). The ecoinvent Database:
446 Overview and Methodological Framework. *Int. J. LCA* 10 (1), 3-9.
- 447 Goedkoop, M., Heijungs, R., Huijbregts, M., De Schryver, A., Struijs, J., Van Zelm, R.,
448 (2013). ReCiPe 2008 – A life cycle impact assessment method which comprises
449 harmonised category indicators at the midpoint and the endpoint level. Ministry VROM
450 Report.
- 451 Hospido, A., Moreira, M. T., Fernández-Couto, M., Feijoo, G., (2004). Environmental
452 Performance of a Municipal Wastewater Treatment Plant. *Int. J. Life Cycle Assess.*, 9
453 (4), 261–271.
- 454 Hospido, A., Moreira, M. T., Feijoo, G., (2008). A Comparison of Municipal Wastewater
455 Treatment Plants for Big Centers of Population in Galicia. *Int. J. Life Cycle Assess.*, 13
456 (1), 57-64.
- 457 ISO 14040, 2006. Environmental Management - Life Cycle Assessment - Principles and
458 Framework: International Standard 14040. International Standards Organisation,
459 Geneva., 2006.
- 460 Lundin, M., Bengtsson, M., Molander, S., (2000). Life Cycle Assessment of Wastewater
461 Systems: Influence of System Boundaries and Scale on Calculated Environmental
462 Loads. *Environ. Sci. Technol.*, 34 (1), 180–186.

- 463 Machado, A. P., Urbano, L., Brito, A. G., Janknecht, P., Salas, J. J., Nogueira, R., (2007). Life
464 cycle assessment of wastewater treatment options for small and decentralized
465 communities. *Water Sci. Technol.*, 56 (3), 15.
- 466 Plastics Europe (2012). *Plastics – the Facts 2012. An analysis of European plastics*
467 *production, demand and waste data for 2011.* Brussels, Belgium.
- 468 Ortiz, M., Raluy, R. G., Serra, L., (2007). Life cycle assessment of water treatment
469 technologies: wastewater and water-reuse in a small town. *Desalination*, 204 (1-3), 121–
470 131.
- 471 Remy, C., Jekel, M., (2008). Sustainable wastewater management: life cycle assessment of
472 conventional and source-separating urban sanitation systems. *Water Sci. Technol.*, 58
473 (8), 1555–62.
- 474 Remy, C. (2010). *Life Cycle Assessment of conventional and source-separation systems for*
475 *urban wastewater management.* PhD Thesis.
- 476 Renou, S., Thomas, J. S., Aoustin, E., Pons, M. N., (2008). Influence of impact assessment
477 methods in wastewater treatment LCA. *J. Clean. Prod.*, 16 (10), 1098–1105.
- 478 Risch, E., Gutierrez, O., Roux, P., Boutin, C., Corominas, Ll., (2015). Life cycle assessment
479 of urban wastewater systems: Quantifying the relative contribution of sewer systems.
480 *Water Res.* 77 (2015), 35-48.
- 481 Rodriguez-Garcia, G., Molinos-Senante, M., Hospido, A., Hernández-Sancho, F., Moreira, M.
482 T., Feijoo, G., (2011). Environmental and economic profile of six typologies of
483 wastewater treatment plants. *Water Res.* 45 (2011) 5997-6010.

- 484 Sansom, M. and Avery, N. 2014. Reuse and recycling rates of UK steel demolition arisings.
485 *Proceedings of the ICE - Engineering Sustainability*, Volume 167, Issue 3.
- 486 Stokes, J. R. and Horvath, A., (2010). Supply-chain environmental effects of wastewater
487 utilites. *Environ. Res. Lett.* 5 (2010), 014015.
- 488 Tillman, A. M., Svingby, M., Lundström, H., (1998). Life Cycle Assessment of Municipal
489 Waste Water Systems. *Int. J. Life Cycle Assess* 3 (3), 145-157.
- 490 Vlasopoulos, N., Memon, F. A., Butler, D., Murphy, R., (2006). Life cycle assessment of
491 wastewater treatment technologies treating petroleum process waters. *Sci. Total*
492 *Environ.*, 367 (1), 58–70.
- 493 Weidema, B. P., Bauer, Ch., Hischier, R., Mutel, Ch., Nemecek, T., Reinhard, J., Vadenbo,
494 C.O., Wernet. G., (2013). The ecoinvent database: Overview and methodology, Data
495 quality guideline for the ecoinvent database version 3. *www.ecoinvent.org*.
- 496
- 497

498 **TABLES**

499 **Table 1:** Description of all the unit processes included in each operation unit studied.

500 **Table 2:** Inventory for the civil works, equipment and operation of the plant divided by unit.

501 **Table 3:** Percentage contribution of civil works (CW), equipment (EQ) and operation (OP)
502 for each operational unit and environmental impact studied.

503

504

505 **FIGURES**

506 **Figure 1:** Procedure followed to obtain the detailed inventory of the WWTP construction.

507 EPD: Environmental Product Declaration.

508 **Figure 2:** Scheme of the Girona WWTP with a separation of each analyzed unit and the
509 inputs and outputs considered for the operational analysis.

510 **Figure 3:** System boundaries considered in this LCA study.

511 **Figure 4:** Contribution of civil works, equipment and operation in each impact category,
512 depending on the life span considered. In each category, the first bar, starting from the left,
513 corresponds to a 10-years life span, the second (middle) bar corresponds to 20 years (base
514 case), and the third (right) corresponds to 30 years.

515 **Figure 5:** Results of the uncertainty analysis for all analyzed categories. In the case of CC
516 category there are two different error bars, the higher uncertainty range includes the
517 uncertainty of N₂O emissions factor provided in Foley et al. (2010).

518 **Figure 6:** Summary of all studies used for the comparison of construction and operational
519 impacts.* Energy consumption as MJ; ** CML estimates eutrophication as a single impact
520 category; *** CML estimates abiotic depletion as a combination of metal depletion (MD) and
521 fossil depletion (FD). The different levels of detail followed in the construction inventories
522 are depicted by the abbreviations

523 **Figure 7:** Cumulative impact from 0% to 100% of the resources consumed in civil works,
524 equipment and operation for CC, HT and MD categories. The contribution of each
525 resource/process from each WWTP unit is shown. Results for other studied impact categories
526 are found in supporting information (Figure S-1 to Figure S-4).

Table 1: Description of all the unit processes included in each operation unit studied.

Operational unit	Elementary processes included
Pumping+pretreatment	Wastewater well reception, pumping station, pretreatment building, part of the connections and part of the unit to dose chemicals
Primary treatment	Primary settlers, units to mix water and chemicals, chamber to measure the flow and part of the connections
Secondary treatment	Biological reactors, secondary settlers, chamber to measure the sludge sent to the sludge line, chamber to measure and pump the sludge sent back to the biological reactor, part of the unit to dose chemicals and part of the connections
Sludge line and deposition	Thickening tanks and buildings for the thickening tanks, chamber to pump the sludge, dewatering building, zone for the dewatered sludge, anaerobic digestion unit and part of connections, final sludge treatment in a composting plant
Others	Chemicals storage, control building, adaptation of the land and sidewalks

Table 2: Inventory for the civil works, equipment and operation of the plant divided by unit.

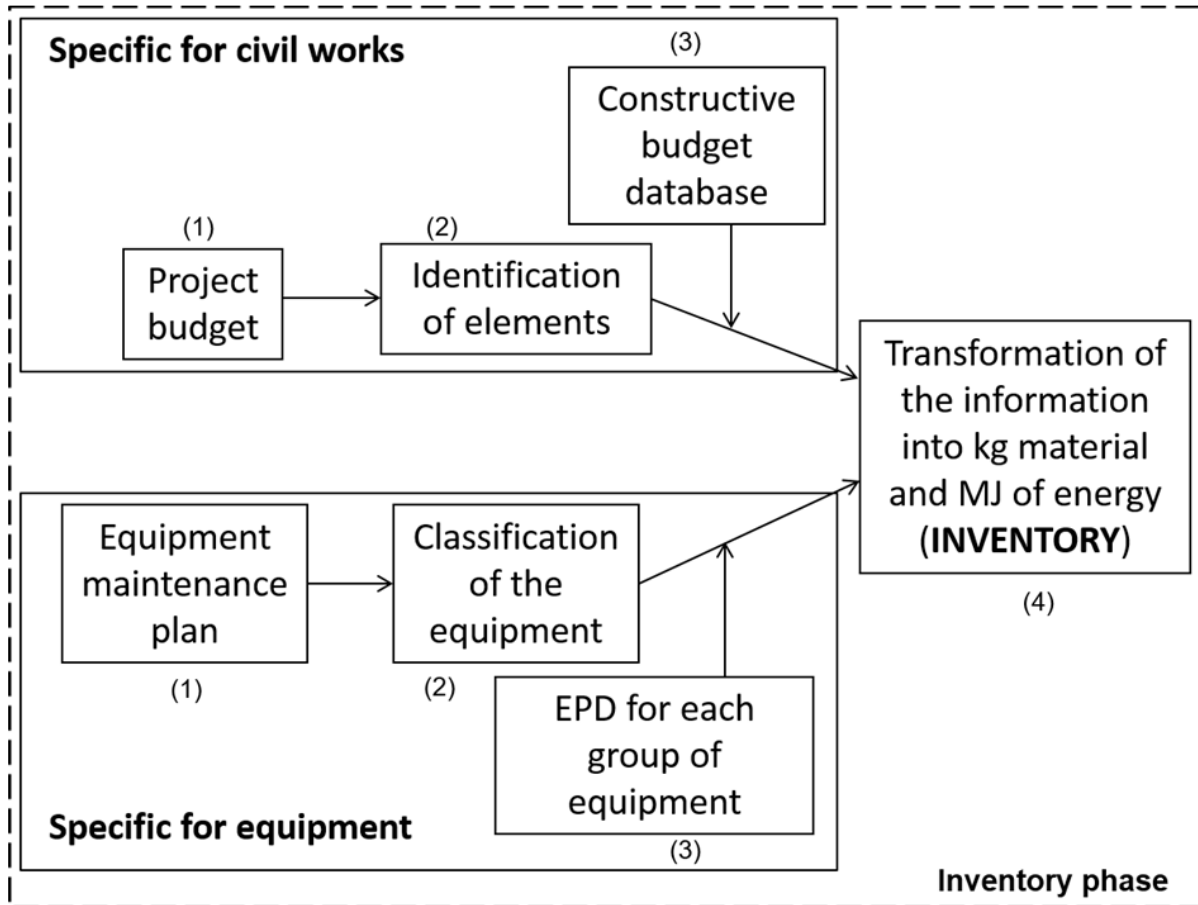
Concept	Unit	Pumping + pretreatment	Primary treatment	Secondary treatment	Sludge line and deposition	Building and services	Full plant
Civil works							
Excavated material	ton·m ⁻³	1.14·10 ⁻⁶	1.00·10 ⁻⁵	2.81·10 ⁻⁴	2.05·10 ⁻⁵	0	3.14·10 ⁻⁴
Energy consumed	MJ·m ⁻³	1.07·10 ⁻³	1.70·10 ⁻³	2.14·10 ⁻²	4.83·10 ⁻³	2.64·10 ⁻³	2.91·10 ⁻²
Total transport	tkm·m ⁻³	7.05·10 ⁻⁴	8.34·10 ⁻⁴	8.89·10 ⁻³	5.33·10 ⁻³	3.56·10 ⁻⁴	1.61·10 ⁻²
Reinforcing steel	kg·m ⁻³	4.96·10 ⁻⁴	4.23·10 ⁻⁴	2.71·10 ⁻³	1.27·10 ⁻³	2.05·10 ⁻⁵	4.92·10 ⁻³
Metal consumption	kg·m ⁻³	3.04·10 ⁻⁵	3.26·10 ⁻⁵	7.52·10 ⁻⁵	5.23·10 ⁻⁵	1.49·10 ⁻⁵	2.05·10 ⁻⁴
Plastic consumption	kg·m ⁻³	2.05·10 ⁻⁴	2.05·10 ⁻⁴	2.12·10 ⁻⁵	5.98·10 ⁻⁴	2.23·10 ⁻⁷	1.22·10 ⁻³
Conglomerates and bricks	kg·m ⁻³	1.28·10 ⁻³	5.58·10 ⁻⁴	7.20·10 ⁻²	2.85·10 ⁻³	6.11·10 ⁻³	8.28·10 ⁻²
Concrete	kg·m ⁻³	1.67·10 ⁻²	1.46·10 ⁻²	7.17·10 ⁻³	2.21·10 ⁻²	1.81·10 ⁻³	6.24·10 ⁻²
Other materials	kg·m ⁻³	7.07·10 ⁻⁵	1.82·10 ⁻⁵	3.79·10 ⁻⁵	2.91·10 ⁻⁵	9.46·10 ⁻⁴	1.10·10 ⁻³
Equipment							
Stainless steel	kg·m ⁻³	6.03·10 ⁻⁵	1.63·10 ⁻⁵	2.86·10 ⁻⁵	3.26·10 ⁻⁵	1.64·10 ⁻⁹	1.38·10 ⁻⁴
Other steel	kg·m ⁻³	8.72·10 ⁻⁶	1.53·10 ⁻⁶	1.09·10 ⁻⁵	1.66·10 ⁻⁵	3.06·10 ⁻⁷	3.79·10 ⁻⁵
Cast iron	kg·m ⁻³	1.94·10 ⁻⁵	3.66·10 ⁻⁶	3.49·10 ⁻⁵	4.66·10 ⁻⁵	-	1.04·10 ⁻⁴
Aluminium	kg·m ⁻³	1.19·10 ⁻⁶	7.25·10 ⁻⁶	2.84·10 ⁻⁶	3.49·10 ⁻⁶	-	8.22·10 ⁻⁶
Copper	kg·m ⁻³	8.29·10 ⁻⁷	1.37·10 ⁻⁷	1.53·10 ⁻⁶	3.89·10 ⁻⁶	-	6.40·10 ⁻⁶
Other metals	kg·m ⁻³	6.18·10 ⁻⁷	1.24·10 ⁻⁶	7.27·10 ⁻⁷	1.53·10 ⁻⁶	-	4.11·10 ⁻⁶
Polypropylene	kg·m ⁻³	1.46·10 ⁻⁷	3.49·10 ⁻⁹	5.90·10 ⁻⁶	3.59·10 ⁻⁸	2.86·10 ⁻⁸	6.10·10 ⁻⁶
Glass fibre reinforced plastic	kg·m ⁻³	2.25·10 ⁻⁶	1.41·10 ⁻⁶	1.42·10 ⁻⁶	2.12·10 ⁻⁶	-	7.20·10 ⁻⁶
Polyethylene	kg·m ⁻³	4.26·10 ⁻⁷	9.24·10 ⁻⁹	4.78·10 ⁻⁷	4.66·10 ⁻⁷	-	1.38·10 ⁻⁶
PVC	kg·m ⁻³	2.67·10 ⁻⁸	1.37·10 ⁻⁷	5.21·10 ⁻⁸	2.59·10 ⁻⁷	-	4.73·10 ⁻⁷
Polystyrene	kg·m ⁻³	-	1.15·10 ⁻⁶	-	-	-	1.15·10 ⁻⁶
Other plastics	kg·m ⁻³	1.39·10 ⁻⁷	1.27·10 ⁻⁷	2.32·10 ⁻⁷	1.17·10 ⁻⁷	8.47·10 ⁻¹⁰	6.15·10 ⁻⁷
Sealing compounds	kg·m ⁻³	4.56·10 ⁻⁷	6.00·10 ⁻⁷	3.24·10 ⁻⁶	2.49·10 ⁻⁶	-	6.77·10 ⁻⁶
Energy	kwh·m ⁻³	3.99·10 ⁻⁵	2.91·10 ⁻⁶	1.06·10 ⁻⁴	4.66·10 ⁻⁵	-	1.95·10 ⁻⁴
Other materials	kg·m ⁻³	7.45·10 ⁻⁷	5.55·10 ⁻⁷	1.22·10 ⁻⁶	9.17·10 ⁻⁷	3.04·10 ⁻⁷	3.74·10 ⁻⁶

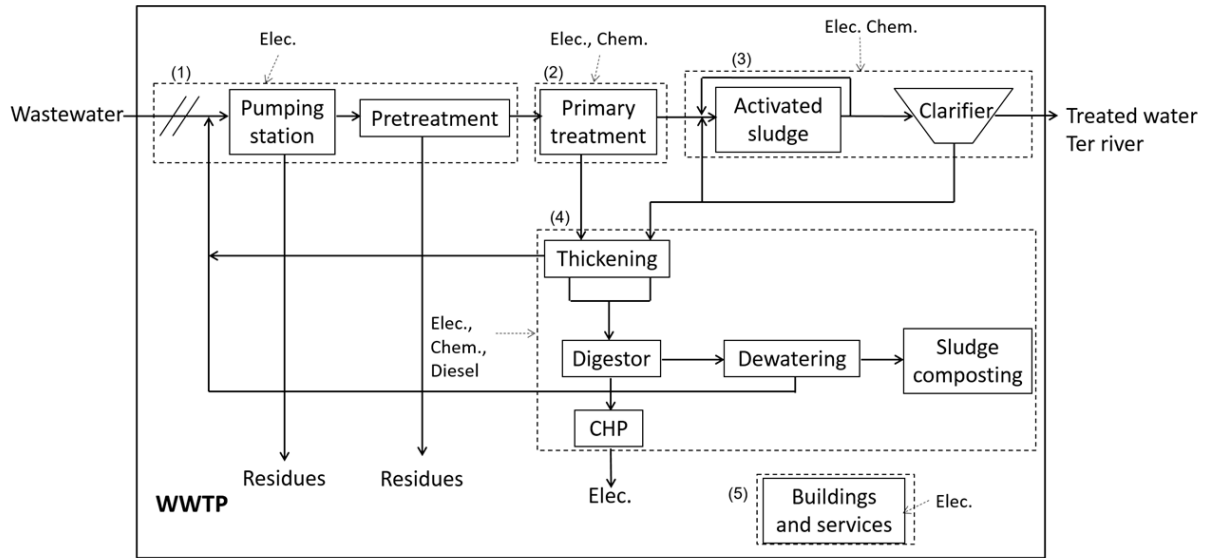
Operation							
Transport	tkm·m ⁻³	1.44·10 ⁻⁵	4.28·10 ⁻⁶	1.44·10 ⁻⁵	1.68·10 ⁻⁵	9.84·10 ⁻⁸	5.01·10 ⁻⁵
Electricity consumption	kwh·m ⁻³	7.19·10 ⁻²	1.75·10 ⁻⁴	2.06·10 ⁻¹	5.03·10 ⁻²	2.12·10 ⁻²	3.50·10 ⁻¹
Electricity production	kwh·m ⁻³	-	-	-	1.31·10 ⁻²	-	1.31·10 ⁻²
Iron chloride	kg·m ⁻³	-	1.29·10 ⁻²	-	-	-	1.29·10 ⁻²
Sodium aluminate	kg·m ⁻³	-	-	4.27·10 ⁻²	-	-	4.27·10 ⁻²
Antifoaming	kg·m ⁻³	-	-	4.74·10 ⁻⁵	2.49·10 ⁻⁴	-	2.96·10 ⁻⁴
Polyelectrolyte	kg·m ⁻³	-	-	-	1.24·10 ⁻³	-	1.24·10 ⁻³
Antioxidant	l·m ⁻³	-	-	-	2.11·10 ⁻⁶	-	2.11·10 ⁻⁶
Diesel	l·m ⁻³	-	-	-	2.73·10 ⁻³	-	2.73·10 ⁻³
Transport	tkm·m ⁻³	1.16·10 ⁻³	3.70·10 ⁻³	1.35·10 ⁻²	1.58·10 ⁻²	-	3.42·10 ⁻²
Direct GHG emissions	kg CO ₂ eq·m ⁻³	-	-	2.17·10 ⁻¹	4.51·10 ⁻²	-	2.62·10 ⁻¹
Residues	kg·m ⁻³	3.74·10 ⁻²	-	-	-	-	3.74·10 ⁻²
Sludge to composting	kg·m ⁻³	-	-	-	7.88·10 ⁻¹	-	7.88·10 ⁻¹
COD emission	kg·m ⁻³	-	-	1.84·10 ⁻²	-	-	1.84·10 ⁻²
NO ³⁻ emission	kg·m ⁻³	-	-	1.76·10 ⁻²	-	-	1.76·10 ⁻²
PO ₄ ³⁻ emission	kg·m ⁻³	-	-	3.32·10 ⁻⁴	-	-	3.32·10 ⁻⁴

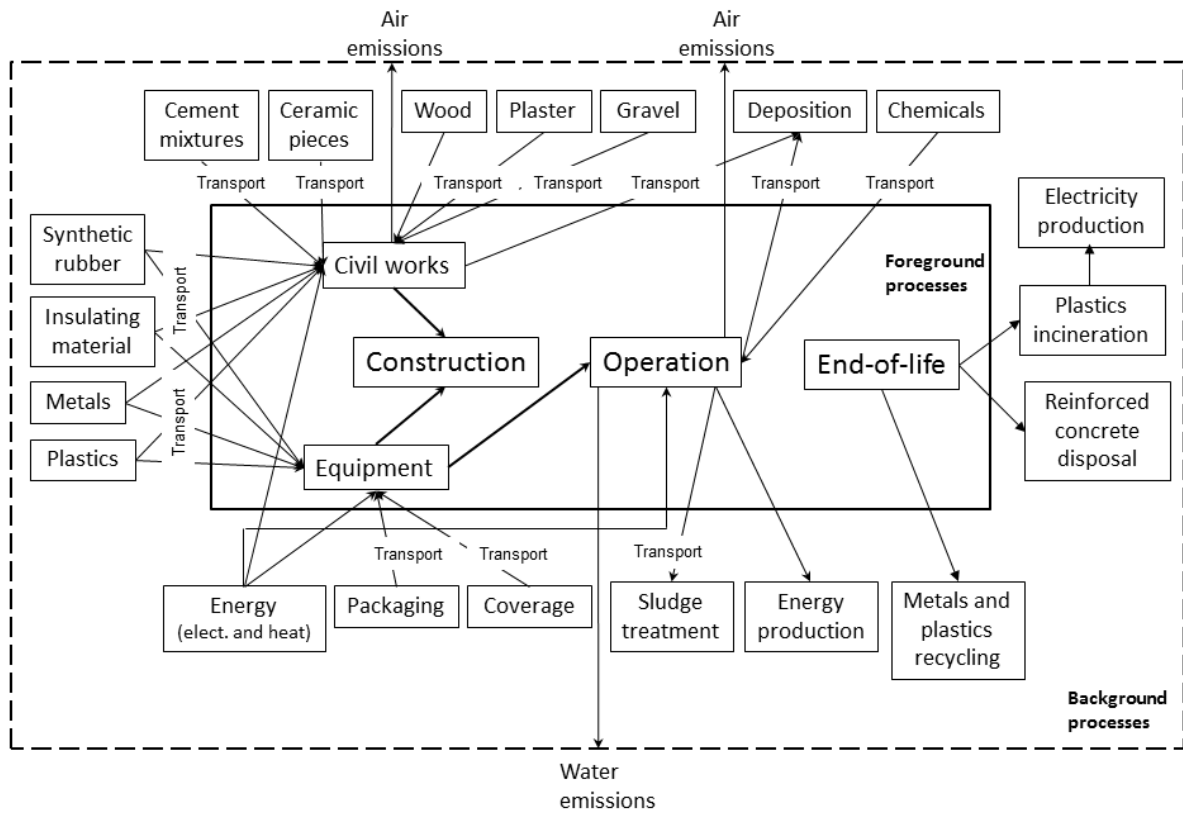
Table 3: Percentage contribution of civil works (CW), equipment (EQ) and operation (OP) for each operational unit and environmental impact studied.

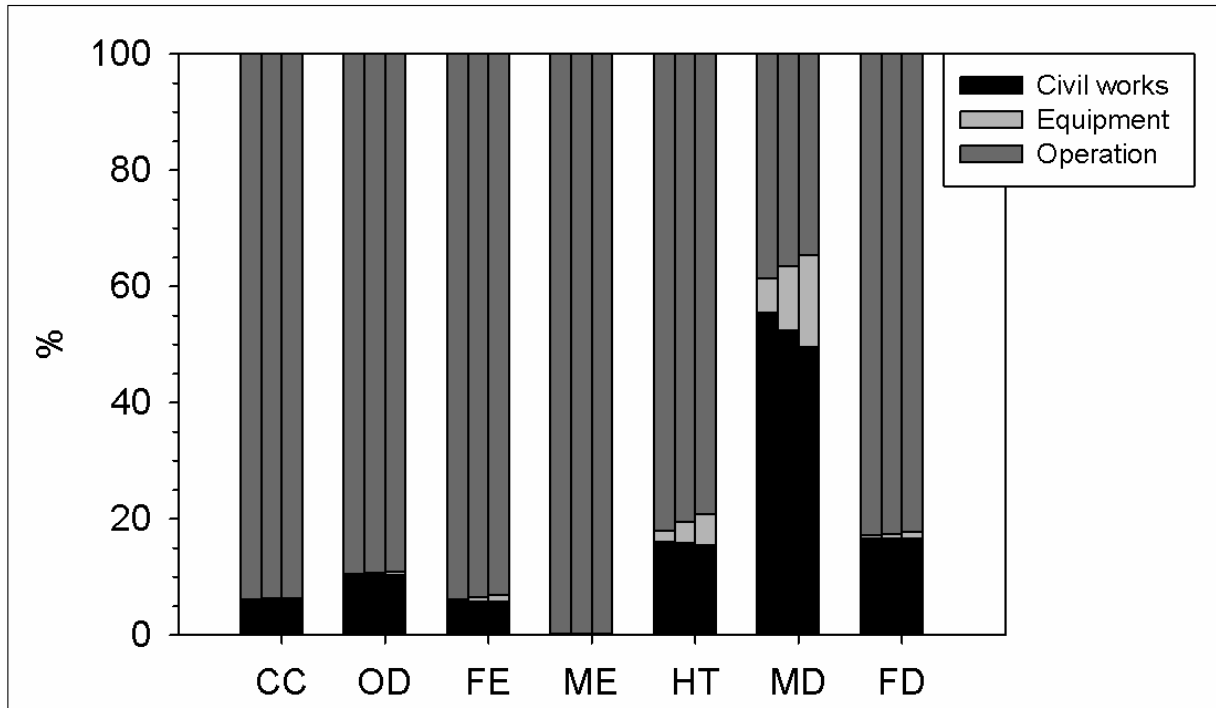
	Pumping + pretreatment			Primary treatment			Secondary treatment			Sludge line			Buildings services and			Total plant		
	CW	EQ	OP	CW	EQ	OP	CW	EQ	OP	CW	EQ	OP	CW	EQ	OP	CW	EQ	OP
CC	8.01	0.78	91.21	20.32	0.73	78.95	5.32	0.09	94.58	5.33	0.20	94.47	9.71	0.02	90.27	6.09	0.20	93.71
OD	9.69	0.88	89.43	5.09	0.32	94.59	7.84	0.18	91.98	111.60	7.58	-19.18 ¹	9.86	0.10	90.04	10.25	0.47	89.28
FE	10.66	2.73	86.62	9.67	1.55	88.79	3.18	0.19	96.63	-104.10	-20.58	224.68 ²	5.74	0.06	94.20	5.75	0.75	93.50
ME	0.86	-0.05	99.19	15.64	0.72	83.64	0.10	0.00	99.90	1.42	-0.44	99.02	11.10	0.03	88.87	0.19	-0.02	99.83
HT	6.67	2.41	90.92	10.17	2.41	87.42	15.07	1.45	83.48	253.52	80.11	-233.64 ¹	9.57	0.10	90.33	16.06	3.29	80.65
MD	41.25	30.41	28.34	25.96	7.57	66.47	58.03	4.68	37.29	80.28	15.35	4.38 ¹	28.18	0.41	71.41	52.83	10.47	36.69
FD	9.78	1.10	89.12	19.85	0.83	79.31	13.59	0.28	86.13	93.52	3.85	2.62 ¹	8.57	0.02	91.41	16.84	0.62	82.53

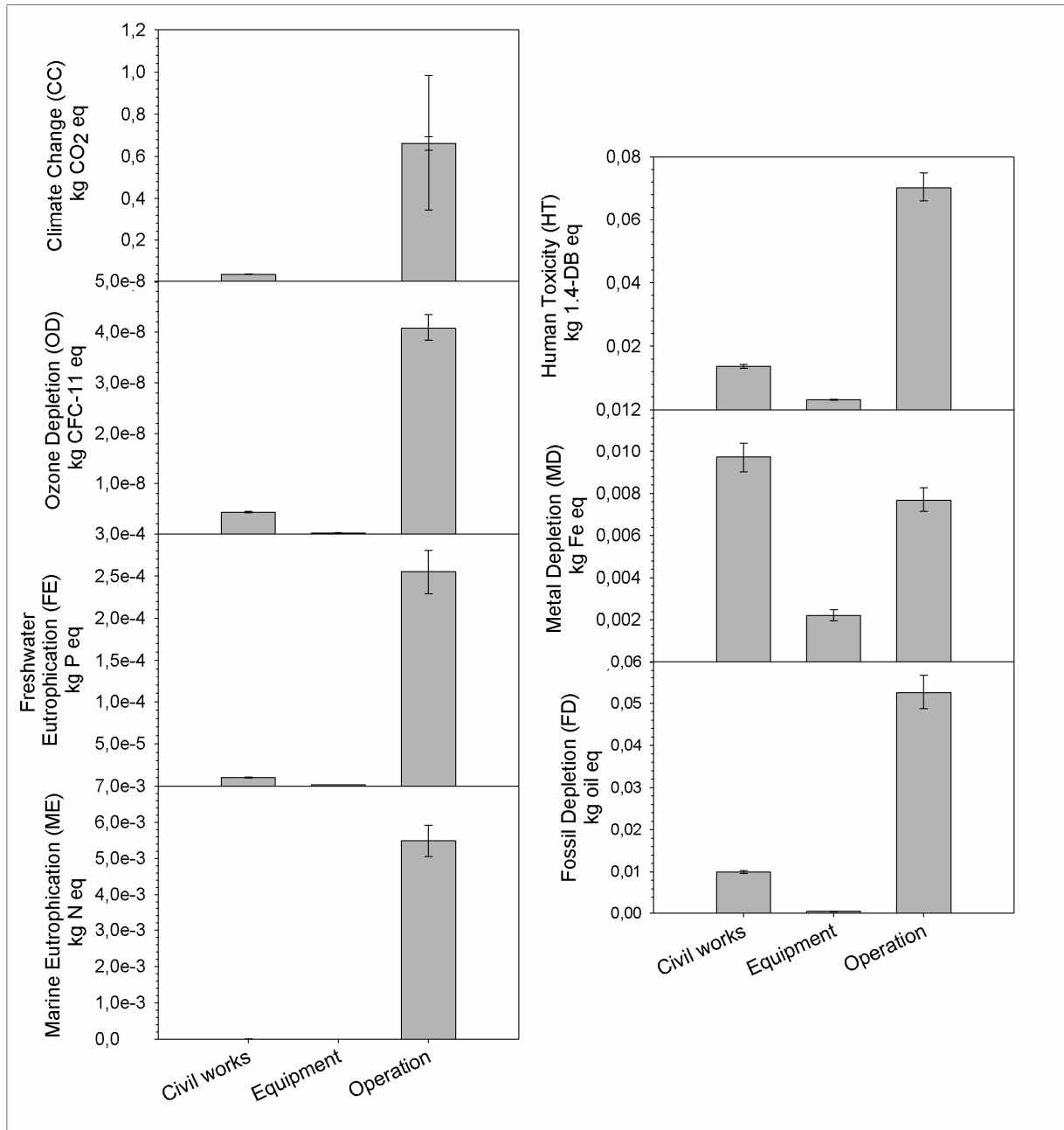
CW (civil works), EQ (equipment), OP (operation), unit of measurement in %. ¹ In these cases, due to the electricity produced during the operation the impact generated by the operation is negative. ² In this case, because the electricity produced is enough to compensate the impact generated by civil works and equipment, for this reason when the contribution of civil works and equipment is compared with the total impact the result is negative.

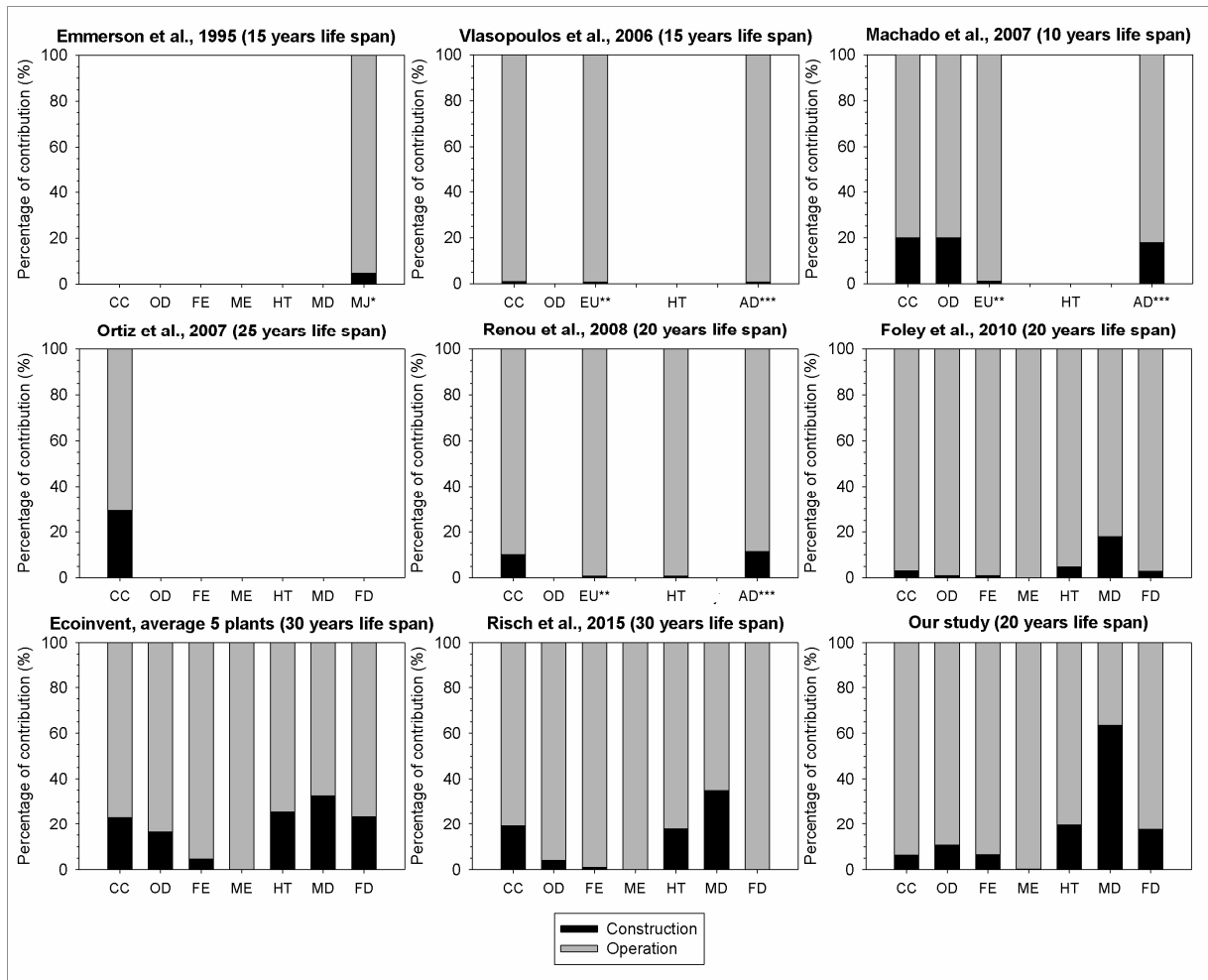




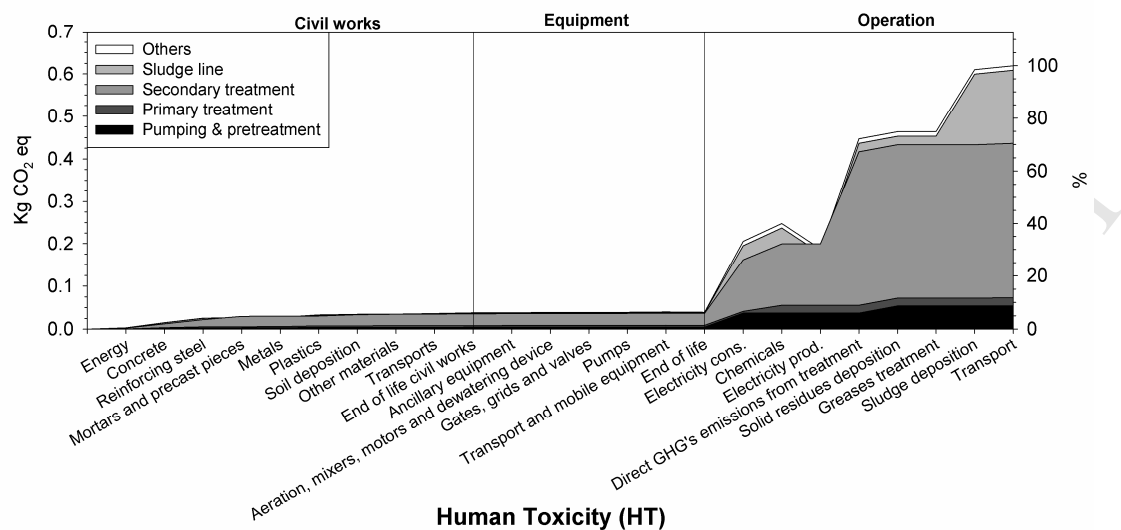




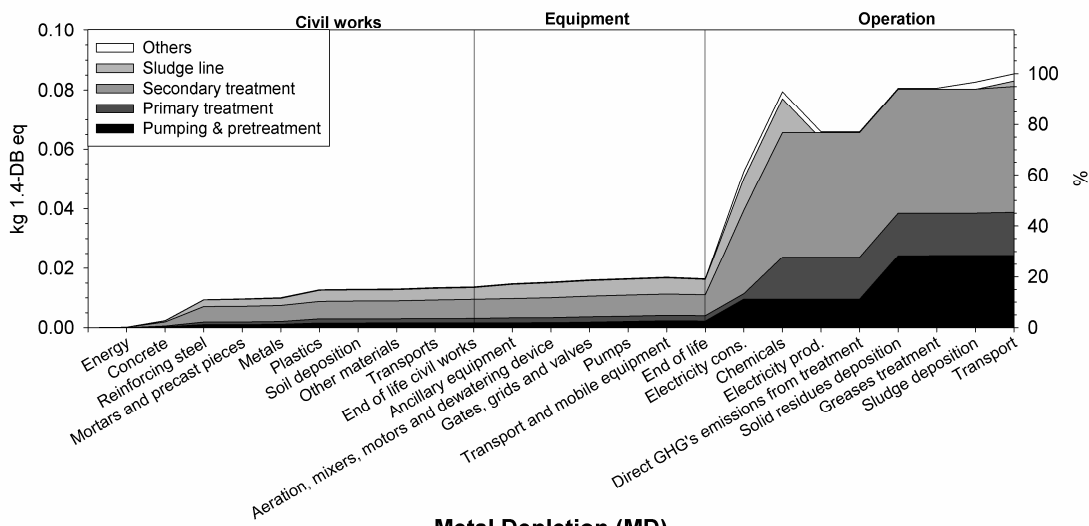




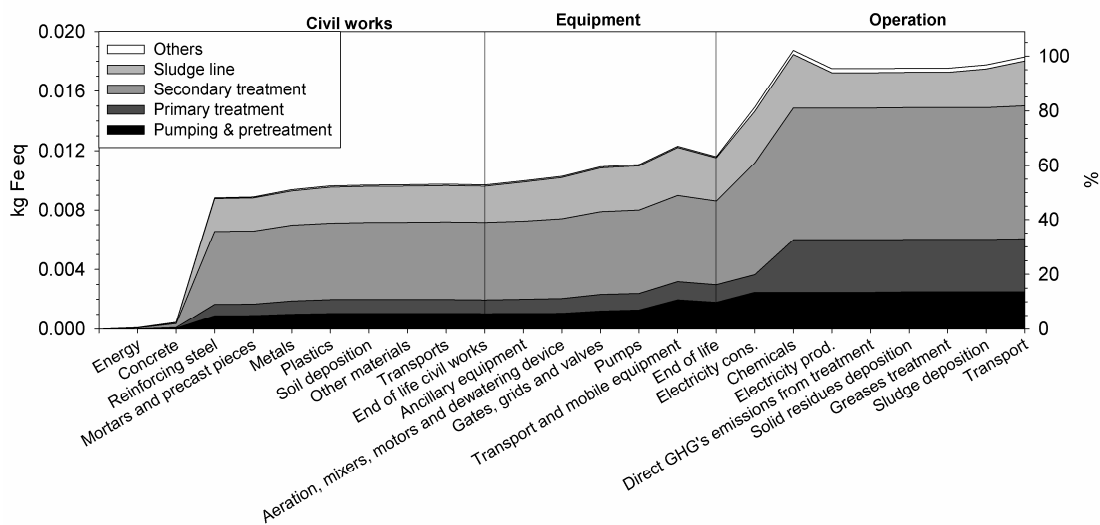
Climate Change (CC)



Human Toxicity (HT)



Metal Depletion (MD)



Highlights

- Detailed civil works and equipment inventories for a large WWTP are provided
- Construction share of the environmental impacts is higher than 5%
- For metal depletion, construction can represent >60% of impacts
- Differences are observed when comparing to Ecoinvent inventories